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# AN IN-DEPTH STUDY OF HEAVY METALS IN SOILS AROUND SELECTED WASTE DUMPSITES IN PARTS OF SOUTHERN NIGERIA USING HEAVY METAL INDICES AND MULTIVARIATE STATISTICAL ANALYSIS

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#### **ABSTRACT**

A detailed study was conducted around waste dumpsite in the study area, to evaluate the levels of heavy metals in the soil. Soil samples were collected at twenty (20) different locations within waste dumpsites for the study. The heavy metals (zinc (Zn), chromium (Cr), copper (Cu), lead (Pb), cadmium (Cd), mercury (Hg), nickel (Ni), arsenic (As), and iron (Fe)) were analyzed using atomic absorption spectrometry (AAS). The resulting data was analysed using principal component analysis (PCA), Pearson correlation, and the heavy metal indices: Potential Ecological Risk (ERI), Geoaccumulation Index (Igeo), Contamination Factor (CF), and Nemerow Pollution Index (PNI). Thesoil contain the heavy metals: zinc (Zn), chromium (Cr), copper Cu, Pb, Cd, Hq, Ni, As, and Fe. Their concentrations, inorder 9.28 to 60.1 mg/kg for Zn, 4.4 to 14.39 mg/kg for Cr, 5.4 to 21.39 mg/kg for Cu, 6.07 to 17.5 mg/kg for Pb, 0.04 to 1.5 mg/kg for Cd, 0.0001 to 0.003 mg/kg for Hg, 7.28 to 20.38 mg/kg for Ni, 0.0001 to 0.003 mg/kg for As, and 1020 to 16540 mg/kg for Fe, The Igeo data indicated that all samples fell within the range of 0 to 2, implying that the sediment quality was largely unpolluted. The PNI yielded a diverse array of results within the study area, PC1 has a loading percentage of 77.8%, PC2 has a loading percentage of 44.4%, PC3 has a loading percentage of 33.3%, and PC4 has a loading percentage of 11.1%. This loading signifies the element's importance in the contribution of pollution within elements and to the soil. Further deductions from soil samples results that was compared to World Health Organization (WHO), and Department of Petroleum Resource (DPR), permissible limit suggested that all samples were below the permissible limit except for Cd. that its concentration was above WHO limit at some locations namely; DO/ 02, 03, 04, 05, 06, 08, 09, 10, 12, 13, and 18, this could be attributed to anthropogenic activities within the study area. These findings can serve as a valuable resource for those responsible for managing dumpsites, aiding in the prevention of heavy metal soil contamination.

KEYWORDS: Risk Assessment, Index, Solid waste, Ecological risk, Statistical techniques.

#### INTRODUCTION:

In recent years, all stakeholders, including ecologists and environmentalists, have found the concentration of heavy metals in soils to be a critical issue (Eyankware, et al., 2024; Osisanya, et al., 2024).

Both geogenic and anthropogenic factors, including metal finishing, paint pigment, livestock wastewater, battery manufacturing, leather tanning, mining activities, municipal waste water sludges, urban composts, pesticides, phosphate fertilizers, vehicular exhaust, energy and fuel production,

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power transmission, automobile mechanic workshops, and chemical lumps from industries and factories, release heavy metals into the environment (Igwe, et al., 2021; Cheng and Zhu, 2007). Heavy metals, such as arsenic (As), iron (Fe), manganese (Mn), cadmium (Cd), chromium (Cr), copper (Cu), lead (Pb), nickel (Ni), and zinc (Zn), particularly contaminate agricultural and industrial soils. Human exposure to certain heavy metals can cause severe ailments such as gastrointestinal, neurological, tumors, head congestion, muscle aches, pharyngeal congestion, eyelid edema, genetic issues, and problems with reproduction (Ogundele, et al., 2020). Currently, soil contamination with heavy metals poses a significant threat to the health of humans, animals, and plants (Omene et al., 2015). It is believed that the heavy metal presence in the soil of the study area primarily stems from garbage dumps (Fig. 1). Globally, human activities generate waste. According to Okove et al. (2022), waste production is an inevitable aspect of life that must be managed by humans. The rapid urbanization in various regions of Nigeria, including Obiakpo and Imo State, is driven by increasing economic and population growth (Eyankware et al., 2022a). Modern agricultural practices, development, industrial activities, and the pursuit of greater crop production have led to substantial releases of heavy metals into the environment (Akinseye, et al., 2023). Soil acts as a primary reservoir for these heavy metals. The extensive application of agrochemicals has heightened the risk of contamination in agricultural areas, significantly affecting the heavy metal levels in the food grown on these lands (Ihedioha, et al., 2019). One of the most difficult tasks facing the government is to provide a healthy and clean environment for the people, particularly in terms of trash management (Eyankware and Ephraim, 2021; Moriasi, et al., 2007). Solid wastes, often referred to as refuse, are discardable materials that outlive their usefulness, while liquid waste, essentially an industrial byproduct, is a major consequence of leachates that contribute to the contamination of water, soil, and air (Awomeso, et al., 2010; Eyankware, et al., 2021a) as shown in Fig. 1a and 1b. The removal of solid waste heaps poses a formidable task for major urban areas, with potential consequences for public health, environmental quality, and the urban landscape (Momodu, et al., 2011; Adedeji and Olayinka, 2013; Walling, et al., 2004; Dimuna, 2004). Waste management is one of those areas that, given the increase in waste output in the urban environment, necessitates additional planning and management. The unprecedented growth of towns and cities has presented numerous obstacles to urban planners, managers, and decisionmakers regarding proper urban development and planning. Large cities frequently have open dumps in undesirable sites, which endanger public health through the spread of diseases and odors.

The uncontrolled disposal of waste has heightened the risk of heavy metal contamination in agricultural lands, adversely affecting the ecosystem and significantly influencing the levels of heavy metals in the food cultivated in these areas. The majority of heavy metals have negative health effects on the kidneys and stomach, even causing cancer (WHO, 1993; Zhao, et al., 2018). Unlike organic contaminants, which microbial action converts to carbon (IV) oxide, heavy metals do not undergo microbial or chemical degradation, and their total concentrations in soil persist for an extended period after introduction (Adriano, 2003; Kirpichtchikova, et al., 2006). Most heavy metals disperse throughout the various minerals of soil, potentially altering their mobility and bioavailability. Studies have shown that the particular physicochemical forms of metals in the environment, as opposed to their total concentration, determine their environmental behavior (Bermond, 2001; Landonin, 2002; Eyankware, 2017; Olajire, et al., 2003). This is essential due to the intricate nature of issues like the emergence of new waste streams. contamination by heavy metals and organic pollutants, along with environmental concerns related to waste processing and treatment, among others (Eyankware et al., 2021b; Elaigwu et al., 2007; Eyankware and Obasi, 2021; Zaini, 2011). Waste dumpsites commonly contain toxic metals, a source of concern and a risk to humans who come into contact with contaminated soils and plants. The chemical compositions of solid waste products frequently cause changes in the physical and chemical features of soil due to contamination (Logan, 2002; Akakuru, et al., 2023; Eyankware, et al., 2023a). One of the most harmful categories of pollutants is heavy metals (Ideriah, et al., 2010; Ulakpaand Eyankware, 2021; Odesa, et al., 2024a). Several studies have assessed the impact of solid waste on the environment globally (Eyankware, et al., 2015; Odesa, et al., 2024b; Onwe, et al., 2024; Igwe, et al., 2021). Various primary index methods, including the geoaccumulation index, ecological risk index, and pollutant load index, are employed to evaluate the ecological risk of soil. The evaluation of soil ecological risk is effectively addressed through the use of indices, which facilitate both qualitative and quantitative assessments of environmental and soil ecological risks. These evaluation techniques can directly reflect the spatially changing trend and pollution level of soil pollutants (Ulakpa, et al., 2016). Many urban environmental pollution sources and intricate risk exposure routes exist. Similarly, study conducted by Thelma et al. (2020) examined the levels of heavy metals and the degree of pollution in soils from abandoned dump sites in Kumasi, Ghana. The findings indicated that moderate levels of heavy metals, including arsenic, cadmium, chromium, copper, lead, mercury, nickel, and zinc, were present in the soils of the dump sites at Amakom and Kronum, with some concentrations surpassing international soil quality standards.

The distribution of these metals was relatively uniform across the study area, although a slightly elevated concentration of lead was observed in the western and northeastern sections of the Kronum and Amakom dump sites, respectively. The index of geo-accumulation and contamination was found to be highest for arsenic at Amakom and for lead at Kronum.Research on heavy metal contamination in soil predominantly emphasizes the quantification of their concentrations in soil samples, often overlooking

their unpredictable nature (Lewandowski et al., 1998; Siepak et al., 1996). While such studies can offer essential insights into the extent of heavy metal pollution, this particular investigation adopts a statistical approach alongside a heavy metal index to assess the impact of solid waste on soil quality. To achieve this, the study set forth several objectives: the evaluation of heavy metals in soil was conducted using various indices, including the potential ecological risk index, the geoaccumulation index (Igeo), the contamination factor, and the Nemerow pollution index.



Fig. 1a. Waste dumpsite around Obiakpu.

# Description of study area, climate, drainage, and topography

The research area is located between latitudes 5°32' N and 5°47' N, and longitudes 6°38' E and 6°59' E, in the southwestern part of Imo State (see Fig. 2). As per the 2006 census, the population of the Imo state stands at 182,500. However, local inhabitants have recently observed a notable increase in population attributed to urbanization and industrialization (NPC, 2006). Considering the ongoing increase in surface temperatures and fluctuations in rainfall, the study area is probably at risk from the impacts of global warming. The study area is located in the wet tropics. The mean annual rainfall ranges from 2250mm to 2500mm.An annual average temperature exceeding 20 °C (68.0 °F) results in a relative humidity level of 75%. During the rainy season, humidity can rise to as high as 90%. This reaches 90 percent in the rainy season. The dry season is characterized by two months of harmattan (from December to late February). Oguta Lake is fed by four rivers: the Njaba, Awbana, Utu, and Orashi (Ahiarakwem et al., 2012), making their recharge crucial. In the rainy season, the Utu Stream, which is perennial, flows directly into the lake, while the Orashi River runs along the southwest side of the lake.



Fig.1b. Waste dumpsite around Obudi.

The Njaba and Awbana Rivers contribute to the lake's water supply throughout the year. According to Mahiarakwem and Onyekuru (2011), the total annual inflow from these rivers and streams is estimated to be 25,801.60 m³. Additionally, the lake is believed to receive 69,000 m³ of return water annually and 138,000 m³ from overland flow (Ifeyinwa and Okorie, 2011). Furthermore, it is estimated that approximately 693,000 m³ of precipitation recharge occurs each year. Ahiarakwem and Onyekuru (2011) also estimate that groundwater contributions to the lake amount to 2,750,400 m³ annually.

#### Geology of the study area

Short and Stauble (1967) categorized the Niger Delta Formations into three distinct groups: Akata, Agbada, and Benin. Within the delta, the Benin Formation consists of a network of aquifers primarily composed of large, highly porous sands and gravels, interspersed with thin layers of shale and clay. Despite the successful drilling of numerous boreholes into the Benin Formation aquifers, many have been abandoned due to high salinity levels (Oteri and Atolagbe 2003). Weber and Daukoru (1976) noted that the Benin Formation, formed during the continental phase of the Niger Delta, has an approximate sediment thickness of 2100 meters.

Onyeagocha (1980) further described the Benin Formation as predominantly consisting of sandstone, sands, and gravel, with clay appearing in lens-like formations. The sands and sandstones vary in grain size from fine to coarse, are partially unconsolidated, and exhibit differing thicknesses. The Benin Formation is predominantly composed of durable freshwater-bearing continental sand and gravel, with

interlayers of clay and shale (Oteri 1990). The Niger Delta Basin, which features extensive porous sands and gravels, contains a multi-aquifer system. In the unconfined aquifer of the coastal beach ridges, a freshwater lens is present above the seawater. In numerous confined aquifers in the study area, freshwater lies beneath saltwater. Certain zones in the Niger Delta have seen saltwater intrusion into limited aquifers, where saltwater-bearing sands rest atop freshwater-bearing sands, as described by Oteri and Atolagbe (2003).

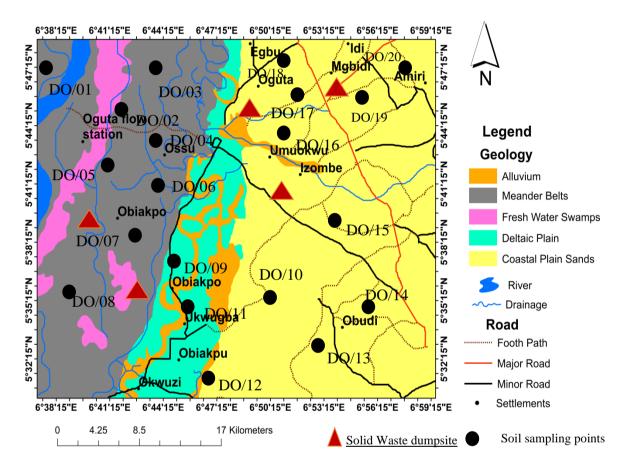


Fig. 2: Geology map of the study area showing soil sample locations

#### **METHODOLOGY**

A total of twenty (20), soil samples were collected around dumpsite within the study area, each sampling point yielded a minimum of three composite soil samples. A steel auger was used to collect soil samples from a depth of 0–15 cm into transparent plastic bags. In the laboratory, samples were air-dried for several days by spreading them out on transparent plastic on a bench. The samples were subsequently passed through a 2 mm sieve and stored in transparent plastic bags until analysis could be conducted. A quantity of 10 g of the sieved air-dried soil samples was transferred into 250 cm³ conical flasks that had been meticulously cleaned, followed by the addition of 100 cm³ of 0.5 M nitric acid. The flasks were then subjected to mechanical shaking for a

minimum duration of 30 minutes using a mechanical shaker. Soil samples are subjected to drying in an oven at temperatures between 105°C and 110°C to facilitate the evaporation of moisture. The materials were then filtered through ashless Whatman filter paper into 100-cm³ plastic bottles. The percentage recoveries were calculated based on the variations observed between the baseline concentrations and those of the spiked samples. For the formulation of solutions intended for cadmium spiking in samples, cadmium nitrate, a widely used reagent with a minimum purity of 99 percent, was employed. Meanwhile, in creating solutions for spiking samples with lead, copper, and zinc, analytical-grade lead nitrate salt along with analytical-grade granules of copper and zinc were utilized.

A reagent blank for each metal was also prepared and processed entirely prior to its application in the sample analysis. Calibration curves were established using analysis-grade metals and metal salts. Using a Varian Techtron AA6 atomic absorption spectrophotometer and associated metallic hollow cathode lamps, the concentrations of cadmium, copper, lead, and zinc were determined. The fuel was acetylene gas, and the support was air. In each case, an oxidizing flame was used. The concentrations of four metals were calculated using calibration curves. To zero the instrument, a reagent blank was employed. The aspiration of standard solutions was conducted first, followed by the aspiration of extracts from soil samples. Prior to their transport to locations with inadequate road conditions for laboratory testing, the soil samples were stored at a station nearer to these roads, which resulted in a delay in sample transmission. The laboratory analyses completed 48 hours after the samples were collected. These measures are crucial for maintaining the integrity of the soil samples.

### Data analysis Soil Pollution Index Calculation

The determination of heavy metal levels in the soil was carried out using the potential ecological risk index, the geoaccumulation index (Igeo) outlined in Tables 4 and 5, along with the contamination factor and the potential ecological risk index provided in Tables 6 and 7.

#### **Potential Ecological Risks Assessment**

$$E_r^i = C_f^i * T_r^i = T_r^i * \frac{C_i}{C_b}$$
(1)

The potential ecological risk index was first proposed by (Hakanson, 1980).

where  $E_r^i$  denotes the potential ecological risk index of metal  $i^{th}$ ;  $T_r^i$  is the toxic response factor of the ith metal. In this study, the  $T_r^i$  of Zn, Cr, Pb, Cu, Ni, and Cd are 1, 2, 5, 5, 5 and 30, respectively which helps assess the potential ecological harm caused by heavy metal contamination by quantifying the ecological risk associated with each metal in the environment. The  $C_f^i$  values of each heavy metal are obtained from (Eq. 4) by using the quotient of concentration of the heavy metal in the soil with the background/target value. To

quantitatively express  $E_r^i$ , five criteria grades were employed:  $E_r^i < 40$ ,  $40 \le E_r^i < 80$ ,  $80 \le E_r^i < 160$ ,  $160 \le E_r^i < 320$  and  $\ge 320$  signifying low, moderate, considerable, high and very high risk, respectively (Hakanson, 1980;Ogunkunle and Fatoba, 2013); Osisanya, et al., 2024). The total of the individual potential ecological risk factors is used to calculate the potential ecological risk index for different heavy metals in soil. The vulnerability of various biological populations is represented, as are any threats posed by heavy metals. All of the detected heavy metals potential ecological risk indices were calculated using (Equation 2)

$$RI = \sum_{i=1}^{n} E_r^i(2)$$

#### (ii). The Geo-accumulationindex (Igeo)

$$I_{geo} = log_2 \binom{c_n}{I.5B_n} (3)$$

As proposed by (Muller, 1979)

Where: Bn is the geochemical background concentration (mg/kg) and Cn is the element n's measured concentration (gg-1)

#### iii. Contamination factor (CF)

The CF was calculated using the (Hakanson, 1980) formula

$$\mathsf{CF} = \frac{c_n}{B_n} \tag{4}$$

Where: Bn is the background/target value, Cn is the metal concentration (Yahaya, et al., 2021)

#### STATISTICAL ANALYSIS

The analysis of the data was conducted utilizing the statistical software SPSS, which facilitated the execution of Principal Component Analysis (PCA).

#### i. Principal Component Analysis (PCA)

PCA is a method for reducing these datasets' dimensionality while enhancing interpretability and reducing information loss (Akakuru et al 2023a). According to Subba Rao et al. (2007), PCA has been a crucial statistical method for figuring out the connections between numerous factors that affect water quality. Component loading in the PCA involved condensing a huge dataset of several variables into a smaller number of linear combinations in the component that accounted for a suitable portion of the total data variance and made it simple to link the variables to the sources or processes using Equation (Eyankware and Akakuru, 2022).

$$\sigma^2 = \frac{1}{N} \sum_{i=1}^{N} (X_{N-}N)^2$$
 Eyankware, et al., 2022b(5)

#### **RESULTS AND DISCUSSION**

Table 1: Concentration (mg/kg) of heavy metals in soil samples within the study area

Sample site	Zn	Cr	Cu	Pb	Cd	Hg	Ni	As	Fe
DO/01	41.10	8.1	11.1	16	0.6	0.00	14.9	0.00	1025
DO/02	46.20	11.2	16.2	17.5	0.7	0.00	13.1	0.00	13540
DO/03	40.1	10.3	20.1	13.5	1.5	0.00	11.4	0.00	12250
D0/04?	32	7.5	12	12.7	1	0.00	13.8	0.00	16540
DO/05	58.2	12	18.2	16.2	0.7	0.00	12.2	0.00	11280
DO/06	60.1	10.4	20.1	15.3	1.3	0.00	14.5	0.01	12440
DO/07	2350	9.03	15.22	11.94	0.2	0.00	20.38	0.02	10493
DO/08	14.24	11.39	10.23	9.39	0.8	0.00	9.2	0.01	10401
DO/09	20.11	7.29	9.29	8.2	1.02	0.01	14.19	0.01	10031
DO/10	11.31	5.39	8.39	14.29	0.74	0.01	18.38	0.02	10344
DO/11	10.2	4.4	5.61	9.64	0.06	0.02	10.33	0.01	14441
D0/12?	9.28	10.22	10.21	15.86	0.75	0.001	9.37	0.01	11031
DO/13	27.29	14.39	14.55	6.07	0.69	0.02	7.28	0.01	10001
DO/14	19.82	11.43	9.29	10.54	0.2	0.01	10.96	0.01	10302
DO/15	11.88	6.49	14.2	9.55	0.31	0.01	13.75	0.01	11192
DO/16	18.19	7.08	11.29	12.65	0.4	0.01	15.65	0.12	1023
D0/17?	20.31	6.95	10.11	13.42	0.21	0.01	9.27	0.001	1022
DO/18	19.93	14.31	5.4	16.54	0.68	0.02	10.11	0.02	1020
DO/19	10.88	10.3	19.15	9.06	0.47	0.01	13.38	0.001	10390
DO/20	20.1	9.82	21.39	14.47	0.04	0.03	8.73	0.01	2030
Min	9.28	4.40	5.40	6.07	0.04	0.00	7.28	0.00	1020
Max	60.1	14.39	21.39	17.50	1.50	0.02	20.38	0.12	16540
Aver	25.85	9.40	13.10	12.64	0.62	0.00075	12.54	0.00065	9501.05

#### **Potential Ecological Risks Assessment**

A Swedish researcher Hakanson, (1980) developed the potential ecological risk index (Table 2). It provides information on the toxicity and ecological sensitivity of contaminant concentrations and was used to evaluate the harmful impacts of toxins on the environment and humans Hakanson, (1980); Weihua; et al., 2010; Suresh, et al., 2012). The tool's original purpose was to be used for evaluating sediment pollution in aquatic environments. Previously, it was utilized to assess the risks associated with soil, dust, and air pollution. (Qingjie, et al., 2008; Qing, et al., 2015; Ogunkunle

and Fatoba, 2013; Iqbal and Shah, 2014; Osipova, et al., 2016). The ERI results presented in Table 3 indicate that the total sample count exceeded 600, signifying elevated ecological risk. This outcome is expected given that there are numerous garbage disposal stations all around the residential neighborhood. The area's anthropogenic activities such waste disposal, and septic tanks, which have made the soil vulnerable to contamination, may be to blame for this extremely high level of ecological danger (Agidi, et al., 2022; Eyankware, et al., 2021a; Onwe, et al., 2022).

Table 2: Ecological Risk Index set limit in Soil (Hakason, 1980).

S/No	Range	Remarks
1	$R_i or E_r^i$	Ecological Pollution Degree
2	$E_r^i < 40 \ or R_i < 150$	Low Ecological Risk
3	$40 \le E_r^i < 80 \text{ or } 150 \le R_i < 300$	Moderate Ecological Risk
4	$80 \le E_r^i < 600 \text{ or } 300 \le R_i < 600$	Considerable Ecological Risk
5	$160 \le E_r^i < 320 \text{ or } 600 \le R_i$	Very high Ecological Risk

Table 3: Results of Potential Ecological Risk (ERI)in soil samples within the study area

Sample	Ti * Pi									ERI
site	Zn	Cr	Cu	Pb	Cd	Hg	Ni	As	Fe	
DO/01	0.293571	0.081	1.541667	0.8	22.5	0.05	0.532143	0.0006	256250	256275.8
DO/02	0.33	0.112	2.25	0.875	26.25	0.01	0.467857	0.0002	338500	338530.3
DO/03	0.286429	0.103	2.791667	0.675	56.25	0.05	0.407143	0.00002	306250	306310.6
D0/04	0.228571	0.075	1.666667	0.635	37.5	0.05	0.492857	0.00002	413500	413540.6
DO/05	0.415714	0.12	2.527778	0.81	26.25	0.15	0.435714	0.0004	282000	282030.7
DO/06	0.429286	0.104	2.791667	0.765	48.75	0.1	0.517857	0.00002	311000	311053.5
DO/07	0.167857	0.0903	2.113889	0.597	7.5	0.02	0.727857	0.00004	262325	262336.2
DO/08	0.101714	0.1139	1.420833	0.4695	30	0.05	0.328571	0.0002	260025	260057.5
DO/09	0.143643	0.0729	1.290278	0.41	38.25	0.05	0.506786	0.00002	250775	250815.7
DO/10	0.080786	0.0539	1.165278	0.7145	27.75	0.005	0.656429	0.00004	258600	258630.4
DO/11	0.072857	0.044	0.779167	0.482	2.25	0.01	0.368929	0.00002	361025	361029
D0/12	0.066286	0.1022	1.418056	0.793	28.125	0.05	0.334643	0.00002	275775	275805.9
DO/13	0.194929	0.1439	2.020833	0.3035	25.875	0.01	0.26	0.00002	250025	250053.8
DO/14	0.141571	0.1143	1.290278	0.527	7.5	0.005	0.391429	0.00002	257550	257560
DO/15	0.084857	0.0649	1.972222	0.4775	11.625	0.05	0.491071	0.00002	279800	279814.8
DO/16	0.129929	0.0708	1.568056	0.6325	15	0.055	0.558929	0.00024	25575	25593.02
D0/17	0.145071	0.0695	1.404167	0.671	7.875	0.005	0.331071	0.00002	25550	25560.5
DO/18	0.142357	0.1431	0.75	0.827	25.5	0.01	0.361071	0.00024	25500	25527.73
DO/19	0.077714	0.103	2.659722	0.453	17.625	0.005	0.477857	0.00002	259750	259771.4
DO/20	0.143571	0.0982	2.970833	0.7235	1.5	0.015	0.311786	0.00042	50750	50755.76
Min	0.066286	0.044	0.75	0.3035	1.5	0.005	0.26	0.00002	25500	25502.93
Max	0.429286	0.1439	2.970833	0.875	56.25	0.15	0.727857	0.0006	413500	413561.5
Aver	0.183836	0.093995	1.819653	0.63205	23.19375	0.0375	0.448	0.00013	237526.3	237552.7

#### **Geoaccumulation Index**

The index of geo-accumulation (Igeo), developed by Muller in 1979 to measures the level of contamination of elemental concentrations in sediment, water, dust, and soil, has been widely used to track these pollutants' pollution status globally (Eyankware and Akakuru, 2022). The classifications of (Igeo) and their respective interpretations are presented in Table 4. The findings from this study's Igeo results (Table 5) reveal that the entire sample is categorized within the range of 0 to 2 (Table 4), suggesting that the soil quality is generally unpolluted, except for iron (Fe), which exhibited significant contamination. Soil

contamination with heavy metals is a significant concern due to its impact on living organisms. Agricultural soils is receiving significant quantities of toxins from various sources. In recent years, rapid urbanization, industrialization, and irresponsible agricultural practices have contributed to the introduction of metal contaminants in farming areas. Heavy metal pollution in agrarian soils might bestow useful issues on soils, impede plant development, and even mischief the strength of people through the defilement of well-established pecking order (Eyankware et al., 2024; Eyankware and Akakuru, 2022).

Table4: Geoaccumulation Index scale (Igwe, et al., 2020; Igwe, et al., 2022; Hazzeman, et al., 2017).

I <sub>geo</sub> value	I <sub>geo</sub> Class	Designation of sediment quality
>5	6	Very highly polluted
4-5	5	Highly populated
>3-4	4	Moderate to highly polluted
2-3	3	Moderately polluted
>1-2	2	Moderated to unpolluted
0-2	1	Unpolluted .
0<	0	Background concentration

Table 5: Results of Igeoin soil samples within the study area

-	Zn	Cr	Cu	Pb	Cd	Hg	Ni	As	Fe
DO/01	0.058916	0.016256	0.061878	0.03211	0.150515	0.001003	0.021359	1.2E-05	51425.96
DO/02	0.066227	0.022477	0.090309	0.03512	0.175601	0.000201	0.018779	4.01E-06	67932.44
DO/03	0.057482	0.020671	0.11205	0.027093	0.376287	0.001003	0.016342	4.01E-07	61460.29
D0/04	0.045871	0.015051	0.066896	0.025487	0.250858	0.001003	0.019782	4.01E-07	82983.94
DO/05	0.083428	0.024082	0.101458	0.032511	0.175601	0.00301	0.017488	8.03E-06	56593.64
DO/06	0.086152	0.020871	0.11205	0.030705	0.326116	0.002007	0.020785	4.01E-07	62413.55
DO/07	0.033687	0.018122	0.084846	0.023962	0.050172	0.000401	0.029214	8.03E-07	52645.13
DO/08	0.020413	0.022858	0.057028	0.018844	0.200687	0.001003	0.013188	4.01E-06	52183.55
DO/09	0.028827	0.01463	0.051788	0.016456	0.255875	0.001003	0.020341	4.01E-07	50327.2
DO/10	0.016213	0.010817	0.046771	0.028678	0.185635	0.0001	0.026347	8.03E-07	51897.57
DO/11	0.014621	0.00883	0.031274	0.019346	0.015051	0.000201	0.014808	4.01E-07	72452.9
D0/12	0.013303	0.02051	0.056917	0.031829	0.188144	0.001003	0.013432	4.01E-07	55344.36
DO/13	0.03912	0.028879	0.081111	0.012182	0.173092	0.000201	0.010436	4.01E-07	50176.68
DO/14	0.028411	0.022938	0.051788	0.021152	0.050172	0.0001	0.015711	4.01E-07	51686.85
DO/15	0.01703	0.013025	0.07916	0.019166	0.077766	0.001003	0.01971	4.01E-07	56152.13
DO/16	0.026075	0.014209	0.062938	0.025387	0.100343	0.001104	0.022434	4.82E-06	5132.561
D0/17	0.029114	0.013948	0.05636	0.026932	0.05268	0.0001	0.013288	4.01E-07	5127.544
DO/18	0.028569	0.028718	0.030103	0.033194	0.170584	0.000201	0.014492	4.82E-06	5117.51
DO/19	0.015596	0.020671	0.106754	0.018182	0.117903	0.0001	0.01918	4.01E-07	52128.36
DO/20	0.028813	0.019707	0.119241	0.029039	0.010034	0.000301	0.012514	8.43E-06	10184.85

#### **Contamination factor (CF)**

In this study, CF was used to calculate the concentration ratio of heavy metals to background values. The values of the contamination factor are described using the following criteria: CF < 1 Low contamination

1≤CF≥3 (CF<1), moderate contamination 3≤CF≥6, significant contamination CF>6, and very high contamination.

(Igwe, et al., 2021; Bhutian, et al., 2017; Eyankware, et al., 2024).

Table 6: Result of the Contamination factor (CF)in soil samples within the study area

Sample site	Zn	Cr	Cu	Pb	Cd	Hg	Ni	As	Fe
DO/01	2.56	0.21	3.17	0.29	0.78	0.01	5.73	0.00	256.25
DO/02	2.88	0.29	4.62	0.31	0.92	0.01	5.03	0.00	338.5
DO/03	2.50	0.27	5.74	0.24	1.97	0.01	4.38	0.22	306.25
D0/04	2	0.19	3.42	0.23	1.31	0.01	5.30	0.22	413.5
DO/05	3.63	0.31	5.2	0.29	0.92	0.01	4.69	0.00	282
DO/06	3.75	0.27	5.74	0.27	1.71	0.01	5.57	0.22	311
DO/07	0.05	0.23	4.34	0.21	0.26	0.00	7.83	0.44	262.325
DO/08	0.89	0.29	2.92	0.17	1.05	0.01	3.53	0.01	260.025
DO/09	1.25	0.19	2.65	0.14	1.34	0.01	5.45	0.22	250.775
DO/10	0.70	0.14	2.39	0.25	0.97	0.05	7.06	0.44	258.6
DO/11	0.63	0.11	1.60	0.17	0.07	0.01	3.97	0.22	361.025
D0/12	0.58	0.26	2.91	0.28	0.98	0.00	3.60	0.22	275.775
DO/13	1.705	0.37	4.15	0.11	0.90	0.00	2.80	0.22	250.025
DO/14	1.23	0.30	2.65	0.19	0.26	5.26	4.21	0.22	257.55
DO/15	0.74	0.17	4.05	0.17	0.40	0.01	5.28	0.22	279.8
DO/16	1.136	0.18	3.22	0.23	0.52	0.01	6.01	0.01	25.575
D0/17	1.269	0.18	2.88	0.244	0.27	0.05	3.56	0.222	25.55
DO/18	1.245	0.37	1.54	0.30	0.89	0.01	3.8	0.01	25.5
DO/19	0.68	0.27	5.47	0.16	0.61	0.02	5.14	0.22	259.75
DO/20	1.25	0.25	6.11	0.26	0.05	0.01	3.35	0.01	50.75
Min	0.05	0.11	1.54	0.11	0.05	0.00	2.8	0.00	25.5
Max	3.75	0.37	6.11	0.31	1.97	5.26	7.83	0.44	413.5
Aver	1.25	0.24	3.74	0.22	0.82	0.48	4.86	0.17	235.88

The sample locations DO/01, 02, 03, 04, 09, 13, 14, and 17 displayed moderate levels of contamination in 55% of instances. Conversely, 30% of the samples (DO/08, 10, 11, 12, 15, and 19) were identified as having "significant contamination," a classification that also applied to moderate zinc contamination at the same rate. Regarding all samples, 55% (DO/01, 02, 03, 05, 06, 07, 13, 15, 16, and 19) were noted for "significant contamination" concerning copper, while 40% (DO/08, 09, 10, 11, 12, 14, 17, and 18) fell into the "moderate contamination" category, and 5% were recorded under "very high contamination." Regarding cadmium, 25% of the analyzed samples fell into the category of "moderate contamination," while a larger proportion, 75%, was identified as having "low contamination." In the case of nickel, a significant 80% of the samples were deemed to have "significant contamination," with 15% categorized as "very high contamination," and 5% as "moderate contamination." For chromium, lead, mercury, and arsenic, every sample was classified as exhibiting "low contamination," whereas all samples pertaining to iron received a designation of "very high contamination.

#### **Pearson correlation**

The correlation matrix is a handy tool for determining how two variables are related. Typically, the correlation coefficient ranges from -1 to +1. The link is referred to as having a negative slope or being anticorrelated if the r-value is close to -1. The relationship is referred to as having a positive slope or being correlated when the value of r is near +1. The points are regarded as uncorrelated if the value is zero (Eyankware and Akakuru, 2022). The findings indicate correlations between the following pairs: Zn and Cu (0.464), Zn and Pb (0.473), Zn and Cd (0.557), and Zn and Hg (0.633). Additionally, correlations were observed between Cr and Ni (-0.449), Pb and As (0.494), and Cd and Hg (0.454), as illustrated in Table 7. These results suggest a weak correlation among the elements, implying that they are largely independent of one another (Akakuru et al., 2021a; Akakuru et al., 2021b).

Table 8: Results of Pearson correlation

	Zn	Cr	Cu	Pb	Cd	Hg	Ni	As	Fe
Zn	1.000								
Cr	.325	1.000							
Cu	.464	.283	1.000						
Pb	.473	.107	.126	1.000					
Cd	.557	.247	.229	.173	1.000				
Hg	.633	.097	.343	.276	.454	1.000			
Ni	069	449	.045	.147	.064	.115	1.000		
As	.358	.176	.125	.494	158	.290	034	1.000	
Fe	.240	119	.127	152	.389	.232	.206	316	1.000

#### **Principal Component Analysis (PCA)**

Principal Component Analysis (PCA) was applied to determine the underlying structure among the heavy metals. Acategorization method known as Principal Component Analysis (PCA) seeks to elucidate the variance among numerous interrelated variables (Eyankware and Akakuru, 2022; Akakuru et al., 2021b). This technique demonstrates relationships between variables, thereby reducing the complexity of the dataset. The covariance matrix of the initial data is used by PCA to extract the eigenvalues and eigenvectors. The uncorrelated (orthogonal) variables known as principal components (PCs) are created by multiplying the initial correlated variables by the eigenvectors (loadings). The transformed observations are referred to as scores: the eigenvalues of the PCs measure their associated variance; and the loadings show the original variables' involvement in the PCs. Each principal component represents a latent factor influencing concentrations across soil samples. PC1 explains the highest variance, indicating it captures the most significant combination of metals. Each principal component represents a latent factor influencing metal concentrations across soil samples. In order to determine the most likely sources of groundwater contamination, it explains the relationship between the variables. Four important principal components were identified. All loadings larger than 0.4 (+ or -) are regarded as important contributors in the analysis interpretation. Zn (0.911), Cr (0.420), Cu (0.585), Pb

(0.553), Cd (0.644), Hg (0.764), and as (0.418) all had positive high loadings in principal component 1 (PC1. 77.7%), according to the results shown in Table 9. This could point to man-made factors like urban or industrial activity, agricultural activities etc. High loading for Cd (0.461), Ni (0.443), As (-0.672), and Fe (0.789) is seen in PC2 (44.4%). could be a reflection of geogenic or anthropogenic sources, such as urban activity and the natural background soil composition. Industrial sources may contribute high concentrations of certain metals like Cd and Ni (positive loadings), while agricultural runoff may reduce others like As (negative loading). PC3 (33.3%): Significant contributions from Cr (-0.656), Pb (0.472), and Ni (0.443) imply localized contamination or an additional source of pollution. PC4 (11.1%): Moderate contributions from various sources and a prominent loading on Cu (0.755) may suggest site-specific or mixed sources. Communalities reflect how much of each variable's variance is explained by the extracted PCs: Higher values (>0.7) suggest that most variance is captured by the PCs. For example, Cu (0.936) and **Zn (0.837)** are well-represented in the PCA model. In summary, according to Table 8's PCA results, PC1 has a loading percentage of 77.8%, PC2 has a loading percentage of 44.4%, PC3 has a loading percentage of 33.3%, and PC4 has a loading percentage of 11.1%. This loading signifies the element's importance in the contribution of pollution within elements and to the soil (Table 9).

Table 9: PCA matrix

	Communalities	Compo	nents		
		PC1	PC2	PC3	PC4
Zn	.837	.911	.001	030	079
Cr	.754	.420	384	656	.005
Cu	.936	.585	.052	145	.755
Pb	.740	.553	372	.472	272
Cd	.796	.644	.461	224	343
Hg	.636	.764	.161	.162	012
Ni	.783	.011	.443	.741	.195
As	.772	.418	672	.380	.041
Fe	.690	.242	.789	081	047
	Eigenvalues	4.55	0.48	0.62	0.24
	Variance (%)	22.14	17.61	10.04	7.35
	Cumulative var. (%)	22.14	39.75	49.79	57.140

## Comaprism of soil result of the study area to International standard

Results from Table 10, revealed all sampled parameters where below WHO, (1981) permissible limit except for Cd. The concentration of Cd soil samples results were below, WHO, (1981) maximum permissible limit except for sample location; DO/ 02, 03, 04, 05, 06, 08, 09, 10, 12, 13, and 18.

According to Agidi et al. (2022) and Li et al. (2015), the concentration of cadmium (Cd) is influenced by the geological parent materials. Cadmium present in the soil arises from a combination of natural processes and human activities. The natural contributors encompass the weathering of geological formations, volcanic eruptions, and the atmospheric deposition of soil particles carried by wind, as well as residues from forest fires (Eyankware, et al., 2024).

On the other hand, human-induced factors, including mining operations, the application of fertilizers—

especially those containing phosphates—and various industrial practices such as combustion and waste management, play a substantial role in elevating cadmium concentrations in the soil (Osisanya, et al., 2024).

Table 10: show soil results obtained from the study area that is compared to WHO, 1981 heavy metal standard for soil, Maximum Permissible Addition (MPA).

S/No	Heavy metal	Target value (mg/kg)	of	soil	Comaprism
1	Zn	50			Below set standard
2	Cr	100			Below set standard
3	Cu	36			Below set standard
4	Pb	85			Below set standard
5	Ni	35			Below set standard
6	As	4.5			Below set standard
7	Co	24			Below set standard
8	Cd	0.8			Above set standard for samples location DO/ 02, 03, 04, 05, 06, 08, 09, 10, 12, 13, and 18

Comaprism of soil result of the study area to (Department of Petroleum Resources) DPR permissible limit.

Findings from Table 11, suggested that soil samples within the study area when compared DPR, (1991) revealed that analyzed soil samples of the study area were below DPR, (1991) set limit, except for Cd, where sample locations DO/03, 04, and 09 where slightly above the set limit. The presence of cadmium

in soil is predominantly due to both natural and human-induced sources. In its natural state, cadmium is present in the Earth's crust as a result of geological weathering and volcanic activity (Eyankware, et al., 2024). In contrast, human activities contributing to cadmium levels include mining, smelting, the irrigation of crops with sewage, as well as the application of phosphate fertilizers and various agricultural practices.

Table 11, show soil results obtained from the study area that is compared to DPR, (199)

S/No	Heavy metal	Target val (mg/kg)	ue of	soil	Comaprism
1	Zn	140			Below set limit
2	Cr	100			Below set limit
3	Cu	36			Below set limit
4	Pb	85			Below set limit
5	Ni	35			Below set limit
6	Cd	0.8			slightly above set limit DO/ DO/ 02, 03, 04,
					05, 06, 08, 09, 10, 12, 13, and 18
7	Fe	47000.0			Below set limit

Source: DPR, (1991)

#### CONCLUSION

Establishing a suitable theoretical framework for analyzing soil conditions across industrial sites, residential neighborhoods, and waste disposal areas is enhanced by calculating indices with diverse attributes. The soil samples were analyzed for the presence of elements such as Zn, Cr, Cu, Pb, Cd, Hg, Ni, As, and Fe. Results from the Igeo test indicated that, apart from Fe, which exhibited significant pollution, all other samples scored between 0 and 2, suggesting that the overall soil quality remains unpolluted. The finding from Pearson correlation reveals a weak association between the elements. Weak correlation also implies that the elements are

generally unrelated to one another. According to the PCA results, PC1 has 77.8% loading, PC2 has 44.4 percent loading, PC3 has 33.3 percent loading, and PC4 has 11.1 percent loading. The primary source of heavy metals in soil within the study area can be attributed to uncontrolled anthropogenic activities. trash dumps. and unsanitary processing environments. Additional analysis of soil sample results, when compared to the permissible limits set by the WHO, indicated that all samples fell below the acceptable threshold, with the exception of cadmium (Cd). The concentration of Cd exceeded the WHO limit at specific locations, namely: DO/02, 03, 04, 05, 06, 08, 09, 10, 12, 13, and 18.

This anomaly may be linked to human activities within the study areaThis study would help alert the inhabitants of the study area to the concentration of heavy metals in the soil within the study area, thereby averting several health-related issues linked to heavy metal contamination. The occurrence of heavy metals in soil could be attributed to anthropogenic sources. It is also advised to conduct regular monitoring of the area for its quality standards be ensured to avoid further declines in soil quality.

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#### Compliance with ethical standards

**Conflict of interest:** The corresponding author certifies that there are no conflicts of interest on behalf of all authors. There are no actual or potential conflicts involving the work.

#### Data availability statement:

Data related to the article are stored in a data repository.

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